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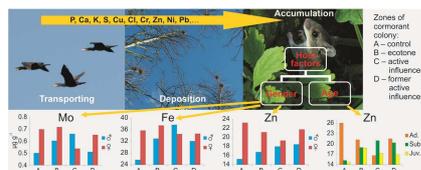
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ACCEPTED MANUSCRIPT

1 **Accumulation of chemical elements in yellow-necked mice under a colony of great**  
2 **cormorants**

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9 **Abstract**

10 This study represents the first investigation into the accumulation of chemical elements in  
11 small mammals inhabiting the territory of a great cormorant colony. Trapping was done in the  
12 Juodkrantė great cormorant colony, one of the largest colonies in Europe. The accumulation of  
13 20 chemical elements in the bodies (muscle and bones) of yellow-necked mice (*Apodemus*  
14 *flavicollis*) was investigated using the energy-dispersive x-ray fluorescence equipment Spectro  
15 Xepos HE. Two groups of positively inter-correlated chemical elements (Mg, Al, P, Ca and Al,  
16 S, Cl, K) were identified. The concentrations of five elements differed significantly between  
17 mice trapped in different zones of the colony with differing intensities of cormorant influence:  
18 the values of K and Cu in *A. flavicollis* increased in line with an increase in the influence of the  
19 cormorants, while the concentrations of Rb and Pb decreased. The concentrations of Mn differed  
20 between zones, but were not related to the intensity of bird influence. Differences in the  
21 concentration of Zn (ANOVA  $F = 24.38$ ;  $p < 0.001$ ), Fe ( $F = 4.60$ ;  $p < 0.05$ ) and Mo ( $F = 4.47$ ;  $p$   
22  $< 0.05$ ) were related to the gender factor, all concentrations being higher in females. The  
23 concentrations of Zn were age-dependent, being highest in adult individuals ( $21.7 \pm 4.5 \mu\text{g g}^{-1}$ )  
24 and exceeding those in subadult ( $19.4 \pm 3.4 \mu\text{g g}^{-1}$ ) individuals or juveniles ( $16.7 \pm 1.3 \mu\text{g g}^{-1}$ ). In  
25 general, the concentrations of accumulated elements in *A. flavicollis* from the territory of the  
26 cormorant colony were lower than in rodents from industrially polluted sites.

27

28 **Keywords:**

29 *Apodemus flavicollis*; *Phalacrocorax carbo*; essential elements; contamination; cormorant  
30 colony.

31

## 32 1. Introduction

33 Atmospheric deposition and anthropogenic activities (agriculture, mining, combustion,  
34 industry) release significant quantities of trace elements and heavy metals. Entering water  
35 ecosystems, these substances accumulate in hydrobionts or deposit in bottom sediments (Yi et  
36 al., 2011; Gajdoš and Janiga, 2015; Hsu et al., 2016). Great cormorants (*Phalacrocorax carbo*  
37 *sinensis*), like other piscivorous birds breeding in colonies, play an important role in transporting  
38 nutrients from water to land ecosystems (Osono et al., 2002; Ellis et al., 2006; Gwiazda et al.,  
39 2010; Klimaszyk et al., 2015; Otero et al., 2015). Most trees, shrubs and other plants in the  
40 territories of cormorant colonies die after few years due to over-fertilization (Garcia et al., 2011)  
41 and are replaced by other plant communities (Ayers et al., 2015; Matulevičiūtė et al., 2018).  
42 Lichen and fungal communities also change (Osono et al., 2002), as do communities of insects,  
43 spiders and lizards (Polis and Hurd, 1996). Most of these changes are related to N and P levels in  
44 the soil, which can be increased by  $10^4$  to  $10^5$  times (Garcia et al., 2011). Trace elements,  
45 including hazardous heavy metals, enter the soil from cormorant excrements (Taraškevičius et  
46 al., 2013). Though some trace elements (Al, Fe, Ni, Cu, Zn, Sr, Mo) are harmful at higher  
47 concentrations, macroelements (Na, Mg, K, Ca) may be beneficial (Pais and Jones, 1997;  
48 Hernout et al., 2016).

49 Concentrations of heavy metals and trace elements have recently been investigated in  
50 different animal taxa, including insects (Aydoğan et al., 2017), crustaceans (Gedik et al., 2017),  
51 fish (Yi et al., 2011; Benzer, 2017), amphibians (Qureshi et al., 2015), reptiles (Nasri 2017),  
52 birds (Kral et al., 2017) and mammals (Lehel et al., 2015; Neila, et al., 2017). Small mammals  
53 have been (Wren, 1986) and remain a favourite object for research into metal and trace element  
54 accumulation (Martiniaková et al., 2012; Gajdoš and Janiga, 2015; Bortey-Sam et al., 2016;

55 Khazaee et al., 2016). Accumulation of heavy metals in small mammals has been well  
56 documented in polluted areas, including near mines (Phelps and McBee, 2009; Bortey-Sam et  
57 al., 2016; Khazaee et al., 2016), power stations (Martiniaková et al., 2010) and paper mills  
58 (Gajdoš and Janiga, 2015). Small mammals serve as suitable objects to study the accumulation of  
59 heavy metals and trace elements, as the concentrations of the metals in the bodies, organs or  
60 tissues of the animals reflect the residues in the soil (Shore and Rattner, 2001; Ieradi et al., 2003;  
61 Martiniaková et al., 2011 and references therein). As a research subject, yellow-necked mouse  
62 (*Apodemus flavicollis*) was chosen for several reasons. In particular, *A. flavicollis* is the most  
63 abundant small mammal species in the territory of the investigated great cormorant colony  
64 (Balčiauskas et al., 2016). Additionally, it is characterized by intensive metabolism, a  
65 granivorous/insectivorous diet and small individual territories (Butet and Delettre, 2011; Gajdoš  
66 and Janiga, 2015). The home range median value for *A. flavicollis* has been identified as 625 m<sup>2</sup>  
67 for males and 551 m<sup>2</sup> for females (Vukićević-Radić et al., 2006). *A. flavicollis* is also known as  
68 proper biomonitor of metal pollution (Petkovšek et al., 2014), with increased levels of metals in  
69 the organism relating to environmental pollution (Martiniaková et al., 2011). In our study, we  
70 analysed chemical elements in the muscles and bones of the skinned bodies of *A. flavicollis*. In  
71 comparison to internal organs, bones accumulate metals over a longer time period (Martiniaková  
72 et al., 2011; 2012) and, as very few *A. flavicollis* individuals live longer than a year, it can be  
73 considered to reflect the elemental load of the year of trapping (Martiniaková et al., 2010; Gajdoš  
74 and Janiga, 2015).

75 Various aspects of the influence of great cormorant and other colonial bird colonies on the  
76 environment are known already (Ayers et al., 2015; Lafferty et al., 2016). However,  
77 investigations into the accumulation of heavy metals and trace elements in small mammals

78 inhabiting these colonies are lacking. The investigated colony in Juodkrantė is one of the biggest  
79 in Europe, with a maximum number of breeding pairs being 3800 in 2015 (V. Knyva, pers. com).  
80 Feeding in both marine and inland waters, including aquaculture, great cormorants are almost  
81 purely piscivorous and the estimated biomass of consumed fish in the Juodkrantė colony is ca.  
82 700 tons per year (Pūtys, 2012), with part of this biomass thereafter reaching the ground in the  
83 colony in the form of excrement and lost fish, as well as dead chicks during the breeding season.

84 This study represents the first investigation into heavy metals and trace elements in small  
85 mammals occupying the territory of a colony of great cormorants. Research concerning the  
86 ecology of mammals in the colonies of great cormorant is still very limited, but decreases in the  
87 diversity of the small mammal community and a reduction in abundance have been described in  
88 relation to the great cormorant colony in Juodkrantė (Balčiauskienė et al., 2014), along with  
89 alterations in the population structure and a decline in body condition (Balčiauskas et al., 2015).  
90 Reductions in *A. flavicollis* body weight, body length and index of body condition, as well as  
91 changes in skull size and shape, were greatest in the most affected zones of the colony  
92 (Balčiauskienė et al., 2015; Balčiauskas et al., 2016). Increased stable isotope signatures in the  
93 small mammals, related to the intensity of cormorant influence, show the consequences of  
94 biological pollution (Balčiauskas et al., 2016). We hypothesized that such changes may be  
95 related to variations in the concentrations of chemical elements in the tissues of the small  
96 mammals. To test the hypothesis, we studied the concentrations of 20 elements in the skinned  
97 bodies of *A. flavicollis* trapped in different zones with different levels of impact by great  
98 cormorants. Differences in accumulation depending on age and gender of mice were evaluated.

99

100

## 101 **2. Material and methods**

### 102 **2.1 Study site**

103 Small mammals were trapped in a colony of great cormorants situated in the western part  
104 of Lithuania near Juodkrantė on the Curonian Spit (WGS 55° 31' 14.22", 21° 6' 37.74"). This  
105 colony is the largest in Lithuania and one of the largest in Europe. It is also distinguished by high  
106 biological pollution reflected even in small mammals, encompassing all investigated aspects of  
107 their biology and ecology (Balčiauskas et al., 2015). The number of breeding pairs reached 3800  
108 in 2015 (V. Knyva, pers. com.) and the area of the colony covers around 12 ha. Four zones with  
109 differing levels of colony influence have been defined in the territory (Fig. 1):

110 Zone A – the control zone. This is outside the colony and there is no direct influence by nesting  
111 cormorants on the habitat.

112 Zone B – the zone of the ecotone. This is located between zones C and D and the surrounding  
113 forests that are not influenced by the colony. There are few nests in this zone and the influence of  
114 the cormorants is weak.

115 Zone C – the zone of active influence. This is the active part of the colony with the highest  
116 concentration of nests and the strongest influence of the colony.

117 Zone D – the zone of former active influence. Nests are already abandoned and trees are dead,  
118 many of them rotten, fallen and decaying.

119

### 120 **2.2 Small mammal sampling**

121 Small mammals were trapped in the middle of September 2015, using snap-trap lines, each  
122 consisting of 25 traps spaced 5 m from each other. Traps were baited with brown bread soaked in  
123 sunflower oil. Exposition of traps was three days (Balčiauskas et al., 2016). Trapping effort was

124 750 trap-days. In total, 132 individuals of six small mammal species were trapped: common  
125 shrew (*Sorex araneus*), bank vole (*Myodes glareolus*), field vole (*Microtus agrestis*), root vole  
126 (*Microtus oeconomus*), harvest mouse (*Micromys minutus*) and yellow-necked mouse  
127 (*Apodemus flavicollis*). The dominant species was *A. flavicollis* (70.5% of all trapped small  
128 mammals) and the subdominant *M. glareolus* (22.7%). Other small mammal species were  
129 insufficiently represented in the various zones of the colony.

130 Chemical analysis was conducted on the dominant species, *A. flavicollis*. Migration of  
131 individuals between zones was investigated in 2013 using live traps and the capture-mark-  
132 recapture method. No migration cases were identified.

133 Before dissection, individuals were weighed (to an accuracy of 0.1 g) and measured with  
134 sliding callipers (accuracy of 0.1 mm). The gender and age of the animals were determined  
135 during dissection. We used three age categories, adult (ad.), subadult (sub.) and juvenile (juv.),  
136 depending on the presence and involution of the *gl. thymus* (involved in adults, disappearing in  
137 subadults, functioning in juveniles) and reproductive status (Balčiauskas et al., 2015). Samples  
138 were placed in separate bags, labelled and stored in a freezer at a temperature below -18 °C.

139

### 140 **2.3 Study of chemical elements**

141 Chemical elements were analysed in 54 individuals (23 males, 31 females / 21 adults, 21  
142 subadults and 12 juveniles) of the dominant species *A. flavicollis*. We used the skinned body  
143 (muscle and bones without intestines, hereafter “body”) to register the presence and  
144 concentration of the following 20 elements: Na, Mg, Al, Si, P, S, Cl, K, Ca, V, Mn, Fe, Ni, Cu,  
145 Zn, Br, Rb, Sr, Mo and Pb. The sampling unit used was body of one individual.

146 Samples were oven dried at 100 °C for 12 hours, crushed in agate mortars and later pre-  
147 mineralized to dry ash at 240 °C to avoid possible ignition and content loss for some volatile  
148 elements (Markova and Rustschev, 1994; Koh et al., 1999). Ashed samples were milled using the  
149 MM 400 mill with zirconium oxide grinding jars and grinding balls (milling time 6 min,  
150 frequency 27 Hz). Milled samples were mixed with the Licowax binder (Fluxana) in the  
151 proportions of 1.25 g of material and 0.28 g of binder (dilution factor 0.816, as recommended by  
152 the equipment manufacturers). Each sample was homogenised and pressed for 3 min using 15  
153 KN (press PP25) to produce 20 mm diameter pellets (Taraškevičius et al., 2017a). The pellets  
154 were analysed by energy-dispersive x-ray fluorescence (EDXRF) equipment Xepos HE (Kleve,  
155 Germany) using TurboQuant (TQ) II for pellets calibration module as elaborated by the  
156 manufacturers. The TQ method combines different procedures: calculation of the mass  
157 attenuation coefficient, using the extended Compton model, and final calibration based on  
158 fundamental parameters method.

159 Samples were re-calibrated using standard bovine muscle (BOVM-1) and the International  
160 Plant-Analytical Exchange (IPE) program. Four extra sub-samples were taken from each of the  
161 IPE Material Samples and from the BOVM-1. Every fifth milled sample (10 extra sub-samples  
162 in total) of *A. flavicollis* was divided into two parts to produce an additional second sub-sample  
163 of the same primary material. The average values of the variation coefficients of paired sub-  
164 samples (RSD) were: < 5% for Na, Mg, Al, P, S, Cl, K, Ca, Mn, Fe, Cu, Br, Rb and Sr; 5–10%  
165 for Ni and Zn; 14% for Mo; and 20–23% for Si, V and Pb. The detection limits ( $\mu\text{g g}^{-1}$ ) of Na,  
166 Mg, Al, Si, P, S, Cl, K, Ca, V, Mn, Fe, Ni, Cu, Zn, Br, Rb, Sr, Mo and Pb were 75, 36, 23, 1.5,  
167 2.2, 0.6, 0.8, 1.2, 0.9, 0.3, 0.2, 0.8, 1.0, 0.5, 0.2, 0.06, 0.06, 0.07, 0.2 and 0.2 respectively.

168 Concentrations of chemical elements were expressed on a wet weight basis in  $\mu\text{g g}^{-1}$  (the weight  
169 loss on average is 412%).

170 Preparations were made at the Nature Research Centre (Vilnius) and the analysis of the  
171 prepared samples was carried out (Spectro Xepos HE) at the Marine Research Institute, Klaipėda  
172 University.

173

#### 174 **2.4 Statistical analysis**

175 In our analysis, we utilized the mean, range (min–max) and standard deviation of  
176 concentrations, Pearson's correlation coefficients and their significance. The normality of the  
177 distribution of concentrations was evaluated using Kolmogorov-Smirnov's test (13 out of 20  
178 elements conformed to normal distribution). Based on conformity to normal distribution,  
179 parametric tests were used. The influences of multifactors were tested using MANOVA. The  
180 influences of zone, gender and age were tested using two-way ANOVA with Wilk's lambda for  
181 significance. The Tukey post-hoc test was used to compare multiple independent groups. The  
182 minimum significance level was set at  $p < 0.05$ . We used STATISTICA 6.0 for Windows.

183

### 184 **3. Results**

#### 185 **3.1 Interaction of host and site factors**

186 The concentrations of chemical elements in *A. flavicollis* depended on which zone of the  
187 great cormorant colony they inhabited (MANOVA Wilks  $\lambda = 0.04$ ,  $F_{3,53} = 2.90$ ,  $p < 0.001$ ) and  
188 the gender of the animals (Wilks  $\lambda = 0.41$ ,  $F_{1,53} = 2.06$ ,  $p < 0.05$ ), but in most cases did not  
189 depend on the age of the animals (Wilks  $\lambda = 0.29$ ,  $F_{2,53} = 1.20$ ,  $p = 0.262$ ). However, animal age  
190 did have a significant impact on the concentrations of Zn in *A. flavicollis*.

191 Differences in the concentrations of chemical elements did not depend upon interaction of  
192 site-based and host factors: age $\times$ site (two-way ANOVA  $F = 1.14$ ,  $p = 0.225$ ) and gender $\times$ site ( $F$   
193  $= 1.16$ ,  $p = 0.264$ ). However, the interaction of two host factors, i.e., age $\times$ gender, did have a  
194 statistically significant influence ( $F = 1.93$ ,  $p < 0.05$ ).

195

### 196 **3.2 Influence of the great cormorants: the zone factor**

197 Depending on which zone of the great cormorant colony the mice had been trapped in,  
198 concentrations of K (ANOVA  $F = 6.45$ ,  $p < 0.001$ ), Mn ( $F = 7.04$ ,  $p < 0.001$ ), Cu ( $F = 3.40$ ,  $p <$   
199  $0.05$ ), Rb ( $F = 14.59$ ,  $p < 0.001$ ) and Pb ( $F = 5.15$ ,  $p < 0.05$ ) differed in the individuals of *A.*  
200 *flavicollis* (Table S1). The concentration of K was significantly higher in mice from zone D than  
201 in zone A (Tukey HSD,  $p < 0.01$ ) and zone C ( $p < 0.01$ ). The concentration of Mn was at its  
202 highest in mice from zones A and C. The concentration of Mn in zone A was significantly higher  
203 than in zone B (Tukey HSD,  $p < 0.01$ ) and zone D ( $p < 0.001$ ), and the concentration of Mn in  
204 zone C was significantly higher than in zone D (Tukey HSD,  $p < 0.05$ ). The concentration of Cu  
205 was at its highest in mice from zone C (significantly higher than in zone A, Tukey HSD,  $p <$   
206  $0.05$ ), though not significantly differing from zone B ( $p = 0.09$ ) and D ( $p = 0.71$ ) (Table S1).

207 The concentration of Rb in mice from zone A was higher than in zones B, C and D (Tukey  
208 HSD, all  $p < 0.001$ ), while the concentration in zone B was higher than in zone D ( $p < 0.05$ ).  
209 Similarly, the concentration of Pb was at its highest in mice from zone A (differences from all  
210 other zones significant at  $p < 0.05$ ). Thus, the concentrations of K and Cu in *A flavicollis*  
211 increased in line with an increased influence of the cormorants, while the concentrations of Rb  
212 and Pb decreased (Table S1, Fig. 2).

213

### 214 3.3 Differences related to the gender and age of the mice: the host factor

215 Differences in the concentrations of Zn (ANOVA  $F = 24.38$ ;  $p < 0.001$ ), Fe ( $F = 4.60$ ;  $p <$   
216  $0.05$ ) and Mo ( $F = 4.47$ ;  $p < 0.05$ ) were related to the gender factor, all concentrations being  
217 higher in females (Table 1). We also found that, for some elements, zone factor had a differing  
218 influence on males and females. In females, the concentrations of K (ANOVA,  $F = 4.35$ ,  $p <$   
219  $0.05$ ) and Cu ( $F = 3.12$ ,  $p < 0.05$ ) differed according to the zone of colony, while this  
220 concentration did not differ in males ( $F = 1.69$ ,  $p = 0.20$  and  $F = 0.73$ ,  $p = 0.16$ , respectively).  
221 Vice versa, the concentrations of Fe significantly differed according to the zone of the colony in  
222 males ( $F = 3.57$ ,  $p < 0.05$ ), but not in females ( $F = 0.58$ ,  $p = 0.64$ ).

223 Age-related differences were found in the concentrations of Zn ( $F = 10.99$ ,  $p < 0.001$ ). Zn  
224 values ( $21.7 \pm 4.5 \mu\text{g g}^{-1}$ ) were highest in the bodies of adult individuals, exceeding those in  
225 subadult individuals (Tukey HSD,  $p < 0.05$ ) and juveniles ( $p < 0.001$ ), while values in subadults  
226 also exceeded juveniles ( $p < 0.05$ ) (Table 1).

227

### 228 3.4 Inter-elemental correlations

229 Two groups of chemical elements that positively and significantly correlated between  
230 each other within the group were identified in the *A. flavicollis* trapped in the territory of the  
231 colony of great cormorants - the first group comprised Mg, Al, P and Ca, while the second group  
232 was Al, S, Cl and K (excluding K-Cl,  $r = 0.252$ ,  $p = 0.066$ ) (Fig. 3). Outside these groups, strong  
233 significant positive correlations in the concentrations were also found between the pairs Br-Fe ( $r$   
234  $= 0.555$ ,  $p < 0.001$ ) and Sr-Pb ( $r = 0.568$ ,  $p < 0.001$ ). Other correlations between the  
235 concentrations of chemical elements are presented in the Table S2.

236

#### 237 4. Discussion

238 Cormorants mediate the transfer of various chemical elements from aquatic to terrestrial  
239 ecosystems. The main source of this transfer is bird excreta (Klimaszyk et al., 2015; Otero et al.,  
240 2015). Pedogeochemical analysis has shown that guano has a low pH, high levels of P, K and Ca  
241 (Breuning-Madsen et al., 2010; Lafferty et al., 2016) and raised concentrations of S, Cl, Cr, Ni,  
242 Cu, Zn and Pb (Taraškevičius et al., 2013). As was emphasized by the latter author,  
243 “geochemical disbalance can be one of the possible reasons of disturbance in natural  
244 ecosystems”. The other possible source of cormorant-borne chemical elements are the fish  
245 brought to the colony as food for chicks, some of these being lost and thus reaching the ground  
246 (Pūtys, 2012). In fish from water bodies in Lithuania, the accumulation of heavy metals follows  
247 the order  $Cd > Pb > Ni > Zn > Cr > Cu$  (Idzelis et al., 2008), with concentrations of Cd and Pb  
248 frequently exceeding the Maximum Tolerable Limit value of both Lithuania and the European  
249 Union (Staniskiėne et al., 2006).

250 In general, the accumulation of trace elements and heavy metals in mammals depends on  
251 habitat, available food, season and host factors, such as species, age and gender (Fritsch et al.,  
252 2010; Lehel et al., 2015; Neila et al., 2017). Although almost any chemical elements can be  
253 detrimental to organisms at high doses, some of these elements (Na, Mg, K, Ca) are not only  
254 essential at lower concentrations, but are also frequently lacking in organisms. Amongst these,  
255 essential elements such as Al, Fe, Ni, Cu, Zn, Sr and Mo can be toxic in high concentrations,  
256 while Pb and Cd may be toxic even at low concentrations (Pais and Jones, 1997; Hernout et al.,  
257 2016).

258 Our study was the first investigation into the accumulation of chemical elements in small  
259 mammals inhabiting the territories of great cormorant colonies. The main factor determining the

260 concentrations of the chemical elements in *A. flavicollis* was the zone of the colony, this  
261 characterizing nest density and bird presence, thus a proxy of bird influence on the local  
262 environment. We found a significant increase in the concentrations of K and Cu and a decrease  
263 in Rb and Pb in *A. flavicollis* trapped in the zones with increased levels of cormorant impact  
264 (Table S1). The overall impact of the influence of cormorants can be considered ambiguous in  
265 terms of advantageous or disadvantageous – the increase in the essential K and Cu is positive, as  
266 is the decrease in the harmful Pb, but the decrease in the essential Rb is negative. A lack of K can  
267 be compensated by heightened Rb intake, extending across all the food web (Nyholm and Tyler,  
268 2000), but a deficiency of Rb is reported as harmful (Gajdoš and Janiga, 2015). In fish at least,  
269 however, excess Rb in combination with heightened Pb, Mo and As may act as a  
270 spermatogenesis inhibitor (Yamaguchi et al., 2007).

271 Concentrations of some trace elements and heavy metals in the bodies of *A. flavicollis* were  
272 gender dependent: females accumulated significantly higher concentrations of Zn, Fe and Mo.  
273 As for the higher concentration of Mo in females, our results confirm those of Gajdoš and Janiga  
274 (2015). Higher concentrations of Zn have also been found in female rats (Bortey-Sam et al.,  
275 2016), as well as human females (Ziola-Frankowska et al., 2015; Taraškevičius et al., 2017b).  
276 Research by Zarrintab and Mirzaei (2017) showed opposite results, there a significantly higher  
277 level of Zn was found in male rats. Out of all the analysed elements, only the Zn concentration in  
278 *A. flavicollis* was age-dependent in the investigated great cormorant colony. The concentration of  
279 Zn increased with age, being highest in adult mice. As well as Zn being involved in the  
280 development of sex organs, it is necessary for normal growth and maturation. Additionally,  
281 juveniles and pregnant or lactating females have increased requirements for zinc (Roohani,  
282 2013). However, it is known that the Zn concentration in mammals is regulated at constant

283 concentrations and is mostly present within a narrow range (Hernout et al., 2016). From this  
284 point of view, our finding of age dependent Zn concentrations in mice from within the cormorant  
285 colony requires further attention. Decreased body size of *M glareolus* and common vole  
286 (*Microtus arvalis*) have been observed in Pb, Fe, Cu and Zn contaminated areas of Slovakia  
287 (Martiniaková et al., 2011), while similar changes in body size and body condition of black-  
288 striped mice (*Apodemus agrarius*), wood mice (*A. sylvaticus*) and greater white-toothed shrews  
289 (*Crocidura russula*) have been observed in other polluted areas (Sánchez-Chardi et al., 2007a,  
290 2007b; Velickovic, 2007). A decrease in body mass in *A. flavicollis* in the expanding part of the  
291 colony typified by fresh nests was also observed in the investigated colony (Balčiauskas et al.,  
292 2015). Comparing heavy metal concentrations in *M. glareolus* from the western part of Lithuania  
293 (Mažeikytė and Balčiauskas, 2003), the average concentrations of Pb ( $0.34 \mu\text{g g}^{-1}$ ) and Cu ( $2.61$   
294  $\mu\text{g g}^{-1}$ ) in the bodies of these voles were higher than those accumulated in *A. flavicollis* from the  
295 great cormorant colony, while the concentration of Ni ( $0.61 \mu\text{g g}^{-1}$ ) did not differ (Table S1).  
296 However, these concentrations are not directly comparable, as *M. glareolus* is known to  
297 accumulate Cd, Pb, Cu and Zn in higher concentrations than *A. flavicollis* (Martiniaková et al.,  
298 2010, 2011). In addition, we have no data on the possible transfer of heavy metals and other  
299 elements with dust, which may have a significant influence on concentrations in small mammals  
300 (Metcheva et al., 2001).

301 Compared to other rodents from industrially polluted sites, the concentrations of the  
302 accumulated elements in *A. flavicollis* from the territory of the cormorant colony were lower.  
303 However, the results are very inconsistent (Table S3). It is known that the accumulation of heavy  
304 metals may differ by up to fivefold between species of shrews, voles and mice in the same  
305 territory (Wijnhoven et al., 2007). We do not discuss these differences with respect to the

306 species, site or organs, but it is clear that biological pollution by cormorants is lower than that by  
307 industrial outputs. Still however, we found differences in some element accumulations that did  
308 depend on the level of the impact of the cormorant colony (Table S1). Concentrations of Pb in *A.*  
309 *flavicollis* from the territory of the great cormorant colony were lower not only than those in  
310 rodents from industrially polluted sites, but also in comparison to concentrations of various  
311 chemical elements in the tissues of the cormorants (Goutner et al., 2011; Misztal-Szkudlinska et  
312 al., 2011). We found two groups of chemical elements in the bodies of *A. flavicollis* with  
313 concentrations correlated within the group, namely Mg, Al, P and Ca, plus Al, S, Cl and K (Fig.  
314 3). Interactions between chemical elements can be related to the specific mineral structure of the  
315 bone tissue and physiological functions of these elements in the organism (Brodziak-Dopierala et  
316 al., 2009). Strong correlations between Mg, P and Ca have been found in human bones (Ziola-  
317 Frankowska et al., 2015), while Gajdoš and Janiga (2015) found strong correlations between S  
318 and K. Likewise in our case, the correlation between S and K was very strong ( $R = 0.808$ ,  $p <$   
319  $0.001$ ). Moreover, we complement this group with correlations with Cl and Al.

320 We may conclude that the accumulation of five out of 20 investigated elements in the  
321 bodies of *A. flavicollis* inhabiting the territory of the great cormorant colony depended on the  
322 intensity of bird influence. However, as identified by previous research, this is hardly likely to be  
323 the sole reason for significant changes in small mammals (Balčiauskienė et al., 2014, 2015;  
324 Balčiauskas et al., 2015, 2016). Many ecological factors are changed due to the biological  
325 pollution of the cormorant colony, including the food base for the small mammals, the  
326 composition of vegetation, the presence of refuges and disturbance by birds. Chemical changes  
327 in the colony resulting from the transfer of materials from the aquatic to terrestrial ecosystem  
328 work in complex with these other changes.

329 The main limitation of our work is the sample size. However, the number of small  
330 mammals inhabiting the colony is finite and can hardly be bigger. Investigations into other  
331 cormorant colonies and other small mammals, such as bank voles (*Myodes glareolus*), will  
332 expand the results of this pilot study and may help to gain a deeper understanding of the  
333 registered chemical changes.

334

### 335 **Conflicts of interest**

336 The authors have no conflicts of interest to declare.

337

### 338 **References**

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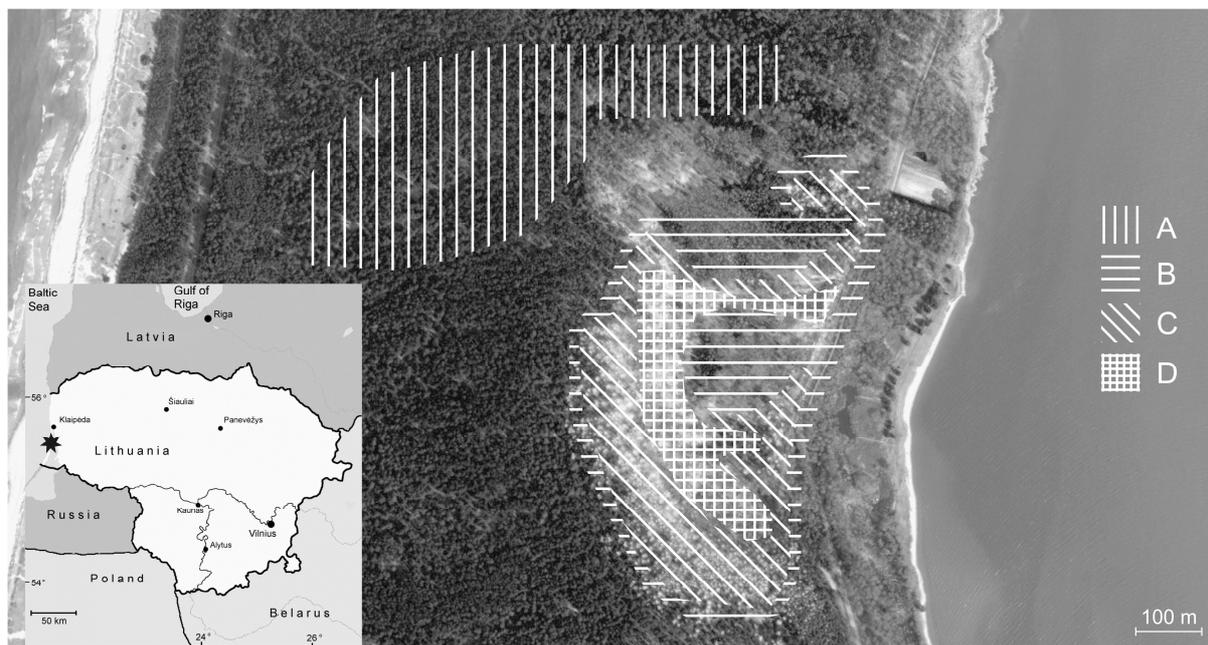
**Table 1.** Concentrations of chemical elements ( $\mu\text{g g}^{-1}$ ) in the bodies of *Apodemus flavicollis* trapped in various zones of the great cormorant colony, according to age and gender groups. Significant differences between groups is presented in bold (ANOVA: \* –  $p < 0.05$ , \*\* –  $p < 0.001$ ). Superscript letters indicate pairwise age-group differences.

Element	Adults (N=21)		Subadults (N=21)		Juveniles (N=12)		Females (N=31)		Males (N=23)	
	Mean $\pm$ SD	Min–Max	Mean $\pm$ SD	Min–Max	Mean $\pm$ SD	Min–Max	Mean $\pm$ SD	Min–Max	Mean $\pm$ SD	Min–Max
Na	708 $\pm$ 103	437–929	681 $\pm$ 86	523–852	688 $\pm$ 69	577–826	706 $\pm$ 71	558–852	676 $\pm$ 108	437–929
Mg	298 $\pm$ 43	199–366	276 $\pm$ 43	207–367	298 $\pm$ 29	255–360	298 $\pm$ 39	227–367	277 $\pm$ 41	199–366
Al	81.7 $\pm$ 9.3	63.9–96.4	79.8 $\pm$ 11.4	59.6–103.9	80.5 $\pm$ 10.6	69.4–100.5	80.9 $\pm$ 10.3	63.2–103.9	80.4 $\pm$ 10.5	59.6–100.5
Si	23.6 $\pm$ 21.5	1.5–86.1	29.4 $\pm$ 36.0	1.5–149.0	29.3 $\pm$ 23.7	2.7–70.1	27.0 $\pm$ 23.9	1.5–82.0	27.2 $\pm$ 33.5	1.5–149.0
P	4952 $\pm$ 528	4106–5996	4772 $\pm$ 564	3869–6099	4949 $\pm$ 322	4420–5469	4971 $\pm$ 496	3973–6099	4760 $\pm$ 501	3869–5996
<b>S</b>	2438 $\pm$ 328	1738–2977	2280 $\pm$ 375	1676–3084	2154 $\pm$ 300	1683–2551	2327 $\pm$ 335	1710–3084	2294 $\pm$ 384	1676–3021
Cl	781 $\pm$ 83	628–941	763 $\pm$ 74	643–932	761 $\pm$ 46	693–831	782 $\pm$ 70	643–941	752 $\pm$ 73	628–894
K	2721 $\pm$ 187	2300–3119	2635 $\pm$ 184	2184–2961	2591 $\pm$ 206	2242–2880	2670 $\pm$ 209	2184–3119	2643 $\pm$ 174	2300–2944
Ca	7483 $\pm$ 1167	5435–9933	6905 $\pm$ 1243	5167–9851	6965 $\pm$ 630	5879–7899	7364 $\pm$ 1071	5167–9851	6845 $\pm$ 1139	5207–9933
V	0.16 $\pm$ 0.02	0.12–0.22	0.15 $\pm$ 0.02	0.11–0.18	0.15 $\pm$ 0.02	0.11–0.18	0.16 $\pm$ 0.02	0.11–0.22	0.15 $\pm$ 0.02	0.11–0.18
Mn	0.78 $\pm$ 0.28	0.48–1.50	0.70 $\pm$ 0.34	0.42–2.03	0.88 $\pm$ 0.39	0.43–1.75	0.82 $\pm$ 0.32	0.43–1.75	0.71 $\pm$ 0.34	0.42–2.03
<b>Fe*</b>	34.7 $\pm$ 4.2	29.9–42.6	33.4 $\pm$ 6.9	22.0–54.9	35.9 $\pm$ 3.8	30.3–42.7	<b>35.8<math>\pm</math>5.5</b>	<b>27.9–54.9</b>	<b>32.7<math>\pm</math>4.8</b>	<b>22.0–42.1</b>
Ni	0.62 $\pm$ 0.10	0.38–0.82	0.61 $\pm$ 0.12	0.43–0.85	0.6 $\pm$ 0.12	0.36–0.78	0.61 $\pm$ 0.11	0.36–0.85	0.6 $\pm$ 0.1	0.43–0.79
Cu	1.66 $\pm$ 0.13	1.37–1.86	1.67 $\pm$ 0.12	1.47–1.95	1.74 $\pm$ 0.16	1.5–2.03	1.69 $\pm$ 0.14	1.46–2.03	1.67 $\pm$ 0.13	1.37–2.00
<b>Zn**</b>	<b>21.7<math>\pm</math>4.5<sup>S,J</sup></b>	<b>15.3–23.2</b>	<b>19.4<math>\pm</math>3.4<sup>A,J</sup></b>	<b>14.5–24.3</b>	<b>16.7<math>\pm</math>1.3<sup>A,S</sup></b>	<b>14.7–18.6</b>	<b>21.3<math>\pm</math>4.3</b>	<b>14.7–28.2</b>	<b>17.5<math>\pm</math>2.3</b>	<b>14.5–23.5</b>
Br	2.38 $\pm$ 0.48	1.28–3.17	2.35 $\pm$ 0.57	0.88–3.59	2.65 $\pm$ 0.37	1.99–3.23	2.41 $\pm$ 0.56	0.88–3.59	2.45 $\pm$ 0.41	1.58–3.11
Rb	6.27 $\pm$ 2.45	3.43–13.49	6.16 $\pm$ 2.25	3.76–12.23	6.18 $\pm$ 1.15	4.74–8.28	6.35 $\pm$ 2.26	3.76–13.49	6 $\pm$ 1.93	3.43–12.23
Sr	3.14 $\pm$ 1.26	1.91–8.03	2.78 $\pm$ 0.51	2.15–4.47	2.60 $\pm$ 0.36	2.11–3.28	3.03 $\pm$ 1.07	2.08–8.03	2.69 $\pm$ 0.49	1.91–3.89
<b>Mo*</b>	0.66 $\pm$ 0.19	0.32–1.09	0.61 $\pm$ 0.17	0.27–0.88	0.57 $\pm$ 0.19	0.24–0.88	<b>0.66<math>\pm</math>0.18</b>	<b>0.32–1.09</b>	<b>0.56<math>\pm</math>0.16</b>	<b>0.24–0.88</b>
Pb	0.05 $\pm$ 0.02	0.02–0.12	0.05 $\pm$ 0.02	0.02–0.08	0.04 $\pm$ 0.01	0.02–0.05	0.05 $\pm$ 0.02	0.02–0.12	0.04 $\pm$ 0.01	0.02–0.08

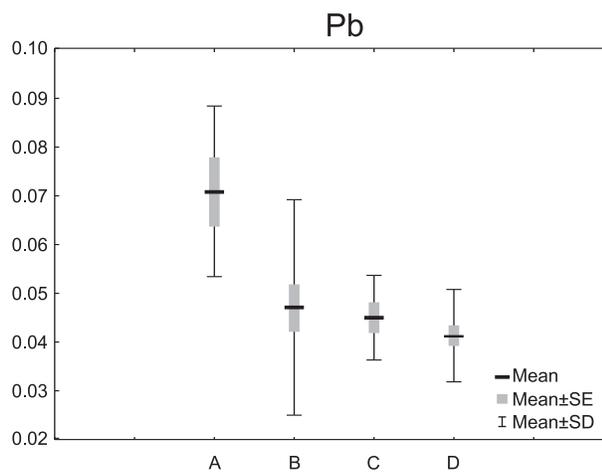
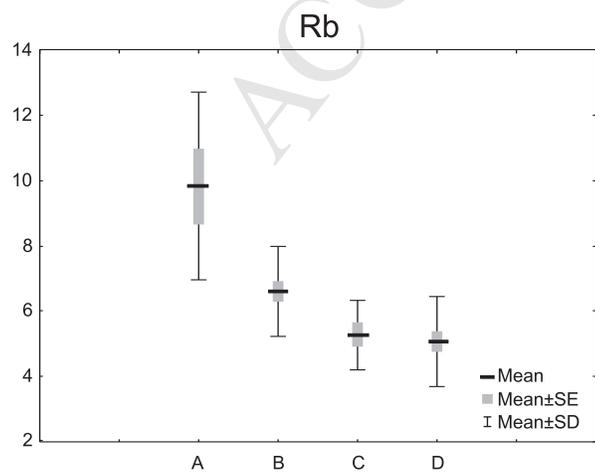
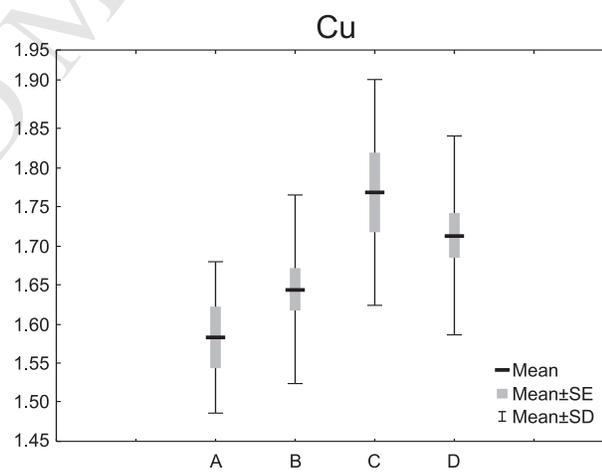
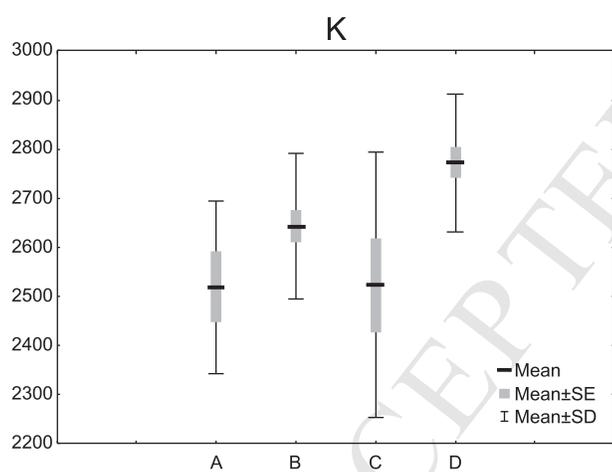
**Fig 1.** Investigation site (marked by star at inlay map of Lithuania) and location of the zones in the great cormorant colony in Juodkrantė, 2015. Zone A – control, zone B – ecotone between the colony and surrounding forest, zone C – active influence of the colony, zone D – zone of the former influence.

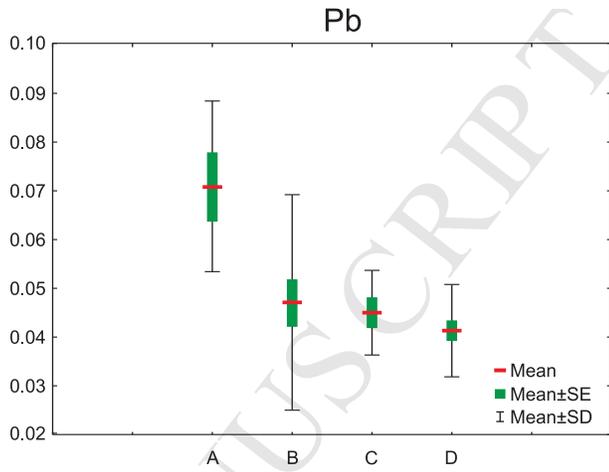
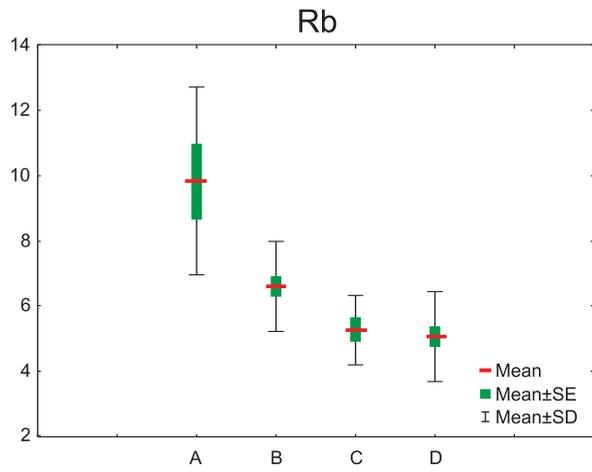
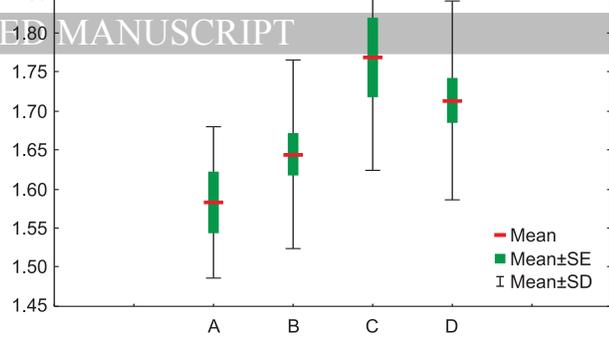
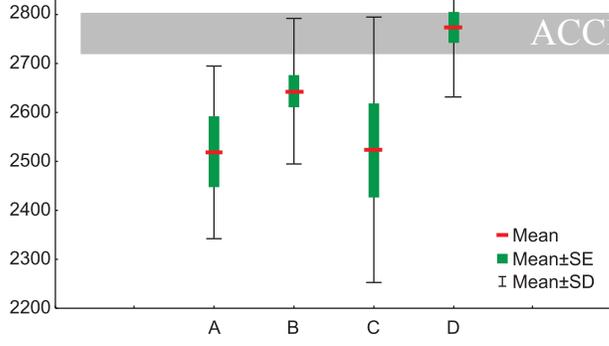
**Fig 2.** Concentrations of K, Cu, Rb and Pb ( $\mu\text{g g}^{-1}$ ) in the bodies of *Apodemus flavicollis* from the different zones of the great cormorant colony (age and gender groups pooled). Zone A – control, zone B – ecotone between colony and surrounding forest, C – zone of active influence of the colony, D – zone of former influence.

**Fig 3.** Intercorrelations between the concentrations of chemical elements in *Apodemus flavicollis* trapped in the territory of the colony of great cormorants (animals from all zones pooled). The correlation between K and Cl ( $r = 0.252$ ,  $p = 0.066$ ) is not shown.

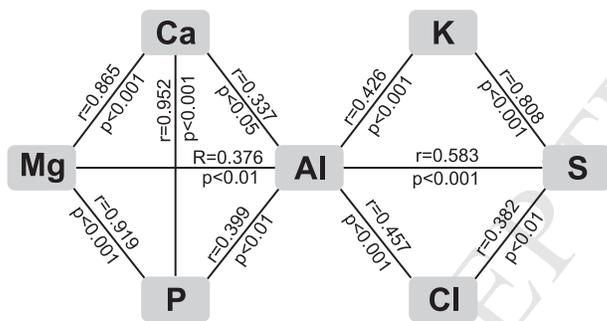


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**Highlights**

- 20 chemical elements in bodies of *Apodemus flavicollis* from a great cormorant colony were studied.
- Concentrations of K, Mn, Cu, Rb, Pb depended on the intensity of cormorant influence.
- Gender-related differences in concentrations of Zn, Fe and Mo were identified.
- Changes in the chemical environment in the cormorant colony affect small mammals.